

ISSN 0046-5070 <http://www.blackwellpublishing.com/fwb>

# *Freshwater Biology*

Edited by Alan G. Hildrew and Colin R. Townsend

Associate Editor: Roger I. Jones, Applied Issues Editor: Richard K. Johnson

Volume 53

Number 11

November 2008



  
WILEY-  
BLACKWELL



# Effects of acidification on the breeding ecology of a stream-dependent songbird, the Louisiana waterthrush (*Seiurus motacilla*)

ROBERT S. MULVIHILL\*, FELICITY L. NEWELL<sup>†</sup> AND STEVEN C. LATTA<sup>‡</sup>

\*Powdermill Avian Research Center, Carnegie Museum of Natural History, Pittsburgh, PA, U.S.A.

<sup>†</sup>School of Environment and Natural Resources, The Ohio State University, Columbus, OH, U.S.A.

<sup>‡</sup>Department of Conservation and Field Research, The National Aviary, Pittsburgh, PA, U.S.A.

## SUMMARY

1. We compared breeding ecology of the Louisiana waterthrush (*Seiurus motacilla*) on acidified and circumneutral streams in the Appalachian Highlands of Southwestern Pennsylvania from 1996 to 2005.
2. Headwater streams impacted by acid mine drainage and/or acidic precipitation showed reduced pH (range 4.5–5.5) compared to four circumneutral streams (pH *c.* 7). Acid-sensitive taxa, including most mayflies (Ephemeroptera), were almost completely absent from acidified streams, whereas several acid-tolerant taxa, especially stonefly (Plecoptera) genera *Leuctra* and *Amphinemura*, were abundant.
3. Louisiana waterthrush breeding density (*c.* 1 territory km<sup>-1</sup>) was significantly reduced on acidified streams compared to circumneutral streams (>2 territories km<sup>-1</sup>). Territories on acidified streams were almost twice as long as on circumneutral streams. Territories usually were contiguous on circumneutral streams, but they were often disjunct on acidified streams. Breeding density declined on one acidified stream that we studied over a 10-year period.
4. Clutch initiation was significantly delayed on acidified streams, on average by 9 days in comparison to circumneutral streams, and first-egg dates were inversely related to breeding density. Birds nesting along acidified streams laid smaller clutches, and nestlings had shorter age-adjusted wing lengths. Stream acidity had no effect on nest success or annual fecundity (fledglings/female). However, the number of young fledged km<sup>-1</sup> was nearly twice as high on circumneutral streams as on acidified streams.
5. Acidified streams were characterized by a younger, less site-faithful breeding population. Individuals were less likely to return multiple years to breed, allowing inexperienced breeders to settle on acidified streams. Pairing success was lower on acidified streams, and we observed four cases of waterthrushes emigrating from territories on acidified streams to nearby circumneutral streams in the following year.
6. We conclude that acidified headwaters constitute lower quality habitat for breeding Louisiana waterthrush. However, breeding birds can apparently compensate for reduced prey resources to fledge young on acidified streams by increasing territory size, foraging in peripheral non-acidified areas, and by provisioning young with novel prey.

*Keywords:* habitat quality, productivity, riparian, *Seiurus motacilla*, stream acidification

---

Correspondence: Robert S. Mulvihill, Powdermill Avian Research Center, 1847 Route 381, Rector, PA 15677, U.S.A.

E-mail: mulvihill@pabirdatlas.org

## Introduction

Acidification of freshwater ecosystems from anthropogenic causes is a major ecological problem worldwide. Reduced pH in streams occurs from several causes, including abandoned mine drainage (Herlihy *et al.*, 1990; Earle & Callaghan, 1998), and acidic precipitation (Schindler, 1988; Herlihy *et al.*, 1993). Drainage from mines is a direct and localized contaminant; in contrast, acid precipitation from rain and snowfall impacts entire regions. But the acidity of a stream may be affected by local differences in soils and bedrock which can neutralize some acid inputs. In addition, streams can be impacted by past acid precipitation events because continued leeching of acidic compounds from the soil can and do affect forest and stream ecosystems for many years (Likens, Driscoll & Buso, 1996; Hames *et al.*, 2002; Kowalik *et al.*, 2007). Estimates suggest that about one-fifth of streams in the northern Appalachian region of the U.S.A. are acidic, with about half of these due to abandoned mine drainage and half from acid precipitation (Herlihy *et al.*, 1990, 1993).

Headwater streams comprise two-thirds of the linear reach of major catchments and, therefore, are critically important to their ecological integrity (Freeman, Pringle & Jackson, 2007). Effects of acidification on wildlife have been found to occur through shifts in trophic relationships (Schreiber & Newman, 1988). Many benthic macroinvertebrate species are acid-intolerant, and changes in species composition with acidification are well documented. Increased acidity shifts the macroinvertebrate community to a comparatively few acid-tolerant taxa with concomitant loss of a large number of acid-sensitive taxa, including most mayfly (Ephemeroptera) species (e.g. Rutt, Weatherley & Ormerod, 1990; Courtney & Clements, 1998; Guerold *et al.*, 2000). Insectivorous birds, which feed at higher trophic levels, can be negatively affected by changes in prey quality and quantity arising from effects of acid waters on macroinvertebrates (Graveland, 1998). Not only prey abundance, but the availability of key, calcium-rich prey, required for egg formation, also is affected (Ormerod *et al.*, 1991).

For example, the distribution and breeding density of the white-throated dipper (*Cinclus cinclus* Linnaeus) in Great Britain is inversely related to

stream pH (Ormerod *et al.*, 1986; Vickery, 1991; Buckton *et al.*, 1998), and acidification has also been found to affect various aspects of dipper breeding ecology, including laying dates, clutch size, territory size and reproductive success (Ormerod *et al.*, 1988, 1991; O'Halloran *et al.*, 1990; Ormerod & Tyler, 1991a; Vickery, 1992). However, some other riparian birds are less affected by stream acidification, presumably because of a lower dependence on aquatic prey (Ormerod & Tyler, 1991b; Vickery, 1991).

The Louisiana waterthrush (*Seiurus motacilla* Vieillot) is the only stream-dependent songbird in eastern North America. This forest-interior Nearctic-Neotropical migrant occupies linear territories along headwater streams where it depends primarily on aquatic macroinvertebrates for food (Robinson, 1995). Waterthrushes have been suggested as an important indicator of the ecological integrity of forested riparian ecosystems (O'Connell *et al.*, 2003; Mattsson & Cooper, 2006). They have been found to occupy streams with a prey biomass containing a higher proportion of mayflies (Ephemeroptera), stoneflies (Plecoptera) and caddisflies (Trichoptera) (hereafter %EPT; Stucker, 2000; Mattsson & Cooper, 2006). The Louisiana waterthrush is considered a species of conservation concern and is listed as a priority species for several Bird Conservation Regions because of its dependence on a specialized habitat threatened by a number of environmental stressors (Rich *et al.*, 2004).

Our objective in this study was to examine how stream acidification affects Louisiana waterthrush population and breeding biology. We compared the ecology and reproductive success of Louisiana waterthrushes breeding on acidified and circumneutral streams over a 3-year period; and on two streams we studied the effects of acidification over a longer, 10-year period. Because ecology is closely linked to resource availability, and given the known negative impact of acid pollution on aquatic macroinvertebrates, we hypothesized that streams with reduced pH would constitute comparatively low quality habitat for breeding Louisiana waterthrush. We predicted that on acidified streams waterthrushes would: (i) breed at lower densities; (ii) have lower pairing and reproductive success and (iii) be less likely to return to the same stream in subsequent years.

## Methods

### Study sites

We monitored 2–3 km reaches of first and second order streams in the Ohio River Drainage in the Laurel Highlands of southwestern Pennsylvania. Our study streams were generally in close proximity (2–30 km apart) to one another. Long-term monitoring from 1996 to 2005 was conducted on two streams, Laurel Run (acidic) and Powdermill Run (circumneutral), both located at Powdermill Nature Reserve (PNR), a field station of the Carnegie Museum of Natural History (Westmoreland County). Six additional streams were monitored from 1998 to 2000, including the acidified (pH 4.5–5.5; Table 1) Linn Run (Westmoreland County), Gary's Run (Somerset County) and Jonathan Run (Fayette County), and the circumneutral (pH *c.* 7.0; Table 1) Camp Run (Westmoreland County), Roaring Run (Somerset County) and Blackberry Run (Fayette County). Passive treatment systems to reduce stream acidification, consisting of a Pyrolucite drain and a Successive Alkalinity Producing System (Gangewere, 1998), were installed to treat two point sources of acidic abandoned mine drainage on Laurel Run in autumn 1997. However, additional diffuse, untreated sources of abandoned mine drainage in the catchment, as well as region-wide acid precipitation, have impeded the recovery of Laurel Run, which remains acidic more than 10 years later.

**Table 1** Measures of water quality at spring base flow (mean  $\pm$  SE) for forested headwater streams in southwestern Pennsylvania

Stream	pH	Alkalinity (mg L <sup>-1</sup> CaCO <sub>3</sub> )	Aluminum (mg L <sup>-1</sup> )
Laurel Run*	4.7 $\pm$ 0.0	0.0 $\pm$ 0.0	7.2 $\pm$ 0.0
Linn Run <sup>†</sup>	5.4 $\pm$ 0.2	2.2 $\pm$ 0.4	0.4 $\pm$ 0.0
Gary's Run <sup>†</sup>	5.0 $\pm$ 0.2	1.4 $\pm$ 0.3	0.4 $\pm$ 0.6
Jonathan Run*	4.5 $\pm$ 0.1	0.8 $\pm$ 0.1	2.1 $\pm$ 0.4
Powdermill Run	7.2 $\pm$ 0.0	18.0 $\pm$ 0.0	<0.1
Camp Run	7.3 $\pm$ 0.1	17.1 $\pm$ 1.5	0.7 $\pm$ 0.6
Roaring Run	7.2 $\pm$ 0.1	12.9 $\pm$ 0.3	<0.1
Blackberry Run	7.1 $\pm$ 0.2	20.6 $\pm$ 5.4	0.1 $\pm$ 0.0

Water samples ( $n = 6$ ) were collected from three points along the length of each stream in 1998–99.

\*Acidification due primarily to point source pollution from acid mine drainage.

<sup>†</sup>Acidification due primarily to acidic precipitation.

All streams were located in contiguous forest (>80% forest cover in a 1 km radius). Mixed deciduous forest surrounding our study streams was characterized by American beech (*Fagus grandifolia* Ehrh.), red maple (*Acer rubrum* L.) and yellow poplar (*Liriodendron tulipifera* L.); some sites also had Eastern hemlock (*Tsuga canadensis* L.), black birch (*Betula lenta* L.) and yellow birch (*Betula alleghaniensis* Britton). Typical forest understorey species present at our sites included common spicebush (*Lindera benzoin* L.), witch hazel (*Hamamelis virginiana* L.) and striped maple (*Acer pensylvanicum* L.). The groundcover was largely dominated by fern species and a diversity of vernal woodland wildflowers.

### Water quality and benthic macroinvertebrates

On each study stream, water samples were collected in June 1998 and 1999 from the lower, middle and upper reaches of the stream during summer base flow for analysis of water chemistry. Results were averaged across years for the length of the study area. Based on water chemistry and known surrounding land use, acidic abandoned mine drainage was the dominant source of chronic acidification for two streams (Laurel Run and Jonathan Run), while the two other acidified streams (Linn Run and Gary Run) showed impacts of episodic acidification from acid deposition. In addition to lower pH levels, aluminum and sulfate levels tended to be elevated at the sites impacted by abandoned mine drainage (Table 1). Circumneutral streams were moderately to well buffered from episodic acidification due to region-wide acid precipitation.

We collected benthic macroinvertebrates from Surber samples in June 1997–99: the two streams at PNR were sampled 1 year; the other study streams were sampled in 2 years. Surber samples were taken from riffles at five stratified random stations along the length of each stream. Individual organisms within each sample were identified to genus, and totals were averaged across stations for each stream (Merritt & Cummins, 1996). Orders represented by >1000 individuals were summed separately, including mayflies, stoneflies, caddisflies and true flies (Diptera). Other macroinvertebrate orders were combined. We calculated the %EPT because the availability of these species has been shown to be correlated with Louisiana waterthrush occupancy (Stucker, 2000; Mattsson & Cooper, 2006).

*Louisiana waterthrush monitoring*

Study areas along each stream were marked with numbered stakes at 50-m intervals, and observations of Louisiana waterthrush were recorded relative to these markers. Waterthrushes on each study area were uniquely marked with a combination of two or three coloured celluloid and one serially numbered U.S. Fish and Wildlife Service ring. Adult waterthrushes were aged as second-year (i.e. <1 year old) or after second-year (>1 year old) based on wing molt limits and rectrix shape (Mulvihill, 1993; Pyle, 1997). Unflattened wing chord and body mass were measured at the time of ringing.

The majority of breeding males (87% across all years) and females (70%) were colour-ringed on our study streams, allowing us to examine population demographics on acidified and circumneutral streams across multiple years. We defined breeding waterthrush each year as either new recruits or returning individuals. By definition, known second-year birds could always be categorized both as new recruits and inexperienced breeders. After the first year of monitoring, we assumed that any other unringed individual arriving on the study stream was also a new recruit, regardless of its age. This included some error since unringed birds could also represent returning individuals that were not successfully ringed the previous year. However, few birds were likely to be erroneously categorized given the high proportion of colour-ringed birds and the observed level of site fidelity of breeding birds. Returning individuals were thus further grouped as either ringed on the same stream the previous year (single return), or ringed even earlier on the same stream (multiple returns). Because of low breeding densities on acidified streams, and because trends appeared to be generally similar between males and females, we pooled the sexes for demographic analysis of populations from acidified and circumneutral study sites.

Waterthrushes were observed on each stream every 1–3 days throughout the nesting period. Territory lengths were determined based on the span of observed upstream and downstream locations of singing and foraging colour-ringed individuals, as well as observed agonistic interactions between males occupying adjacent territories. Males that continued to sing throughout the breeding season, or that were

never observed with a female or at a nest, were considered to be unpaired.

We attempted to document all nesting attempts of every waterthrush pair in our study areas through systematic searching of likely nesting sites and observation of adult waterthrush for behavioural cues. In our study reaches, based on the number of paired waterthrushes on territories, an estimated 90% of all nests were found. These were most often found during the nest building or incubation stages. Clutch initiation dates, if not directly known, were backdated from observed hatching or fledging dates, assuming a 13-day incubation period (beginning with the last egg laid) and a 10-day nestling period (Robinson, 1995). Nest contents were subsequently checked every 3–4 days to monitor egg and chick survival. Many nests were under almost daily observation from a blind for a related study of parental provisioning rates and nestling diet. Nestlings were uniquely colour-ringed when they were 7 to 10 days old. If hatching was not directly observed, the ages of nestlings were estimated based on laying dates, by backdating from fledging dates or based on developmental (e.g. feathering) stage (Eaton, 1958). For analysis, individual nestling measurements were standardized for age by subtracting the mean of the age at which the measurement was taken and dividing by the standard deviation for that age. With a mean of zero and standard deviation of one for each age, individual measurements were on the same scale as a relative difference from other individuals of the same age. Measurements could then be compared across ages (McCarty, 2001). Observations of ringed fledglings and/or of adults carrying food were used to confirm nesting success; nests were considered successful if at least one young fledged.

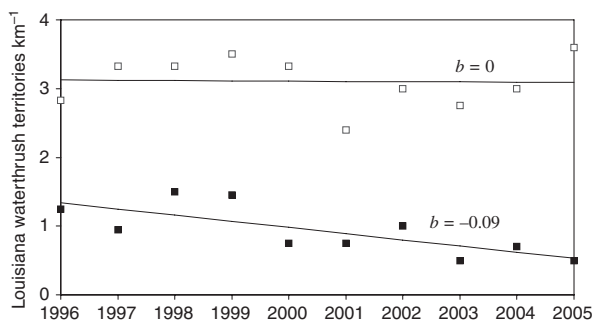
*Statistical analysis*

The software package R for Windows version 2.6.1 (R Development Core Team, 2007) was used to perform statistical tests. Data were tested for normality using Shapiro–Wilk normality tests and data were log-transformed when appropriate. Two-tailed *t*-tests with unequal variance were used to compare between groups. Linear regression was used to examine relationships between variables. Residual plots of the regression analysis were examined. Data presented are means  $\pm$  SE values unless otherwise stated.

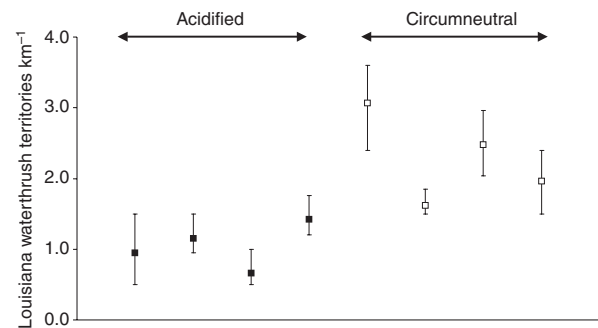
For count data we used a  $2 \times 4$  contingency table with Fisher's exact test to compare between groups. Nest success was calculated according to the Mayfield method (Mayfield, 1975; Johnson, 1979). Differences between stream types were compared using a chi-squared test in the programme CONTRAST (Hines & Sauer, 1989). For all tests, we interpreted a  $P \leq 0.05$  as representing a statistically significant difference or change;  $0.10 > P > 0.05$  was interpreted as a marginally significant tendency.

## Results

We monitored a total of 18 km on eight study streams. We identified 55 territories on acidified streams and 152 territories on circumneutral streams. From 1996 to 2005, Louisiana waterthrush breeding density was significantly lower on acidified Laurel Run than circumneutral Powdermill Run ( $t = -13.12$ ,  $n = 20$ ,  $P < 0.01$ ). Density declined significantly on the acidified stream ( $r^2 = 0.56$ ,  $P = 0.01$ ; Fig. 1), but did not decline on the circumneutral stream ( $r^2 = 0.00$ ,  $n = 10$ ,  $P = 0.92$ ). From 1998 to 2000 breeding density on all acidified streams combined was lower, averaging  $1.0 \pm 0.1$  territory  $\text{km}^{-1}$  (range: 0.5–1.8), compared to  $2.6 \pm 0.2$  territories  $\text{km}^{-1}$  (range: 1.5–3.6) on circumneutral streams (1998:  $t = -3.08$ ,  $n = 8$ ,  $P = 0.02$ , 1999:  $t = -3.00$ ,  $n = 8$ ,  $P = 0.03$  2000:  $t = -2.89$ ,  $n = 8$ ,  $P = 0.04$ ) (Fig. 2). However, density of territories was variable on circumneutral streams, and there was some overlap in breeding density, with two streams (Camp and Blackberry) supporting  $<2.0$  territories  $\text{km}^{-1}$  at least in some years. Territories for which



**Fig. 1** During a 10-year period Louisiana waterthrush breeding density remained consistently lower on acidified Laurel Run (solid squares) compared to circumneutral Powdermill Run (open squares). Density of breeding birds declined significantly on the acidified stream during this period.



**Fig. 2** Mean ( $\pm$ SE) breeding density of Louisiana waterthrush on acidified (solid squares) and circumneutral streams (open squares) in southwestern Pennsylvania. Laurel Run and Powdermill Run were monitored 1996–2005; other streams were monitored 1998–2000.

approximate length was available averaged  $630 \pm 50$  m ( $n = 32$ ) on acidified streams, and  $370 \pm 10$  m ( $n = 88$ ) on circumneutral streams, and this was significantly different ( $t = 4.56$ ,  $n = 128$   $P < 0.01$ ). Territories were as small as 250 m on both stream types, but were  $\geq 700$  m only on acidified streams. Extensive areas were unoccupied on all acidified streams, but such occupancy gaps were rare on circumneutral streams.

Mean abundance of all macroinvertebrates on acidified streams was  $184 \pm 32$  individuals  $\text{m}^{-2}$  and was not significantly different from  $195 \pm 29$  individuals  $\text{m}^{-2}$  on circumneutral streams ( $t = -1.03$ ,  $n = 70$ ,  $P = 0.30$ ). Mean biomass on acidified streams ( $62 \pm 13$  mg  $\text{m}^{-2}$ ) was significantly less than on circumneutral streams ( $124 \pm 21$  mg  $\text{m}^{-2}$ ;  $t = -2.52$ ,  $n = 60$ ,  $P = 0.01$ ). Mayflies as a component of the macroinvertebrate community almost completely disappeared from acidified streams, whereas two stonefly genera, *Leuctra* and *Amphinemura* were especially abundant (Fig. 3). As a result, the overall %EPT did not change significantly ( $t = 0.91$ ,  $n = 15$ ,  $P = 0.37$ ). Louisiana waterthrush breeding density was not related to overall %EPT ( $r^2 = 0.00$ ,  $n = 21$ ,  $P = 0.97$ ), but there was a significant relationship between waterthrush breeding density and %EPT when *Leuctra* and *Amphinemura* stoneflies were excluded from the analysis ( $r^2 = 0.68$ ,  $n = 21$   $P < 0.01$ ; Fig. 4).

Laying commenced as early as 18 April during our study, and the date of laying in the first nest on each stream was significantly related to annual breeding density ( $r^2 = 0.47$ ,  $n = 37$ ,  $P < 0.01$ ; Fig. 5). On circumneutral streams, laying in the earliest nest each year



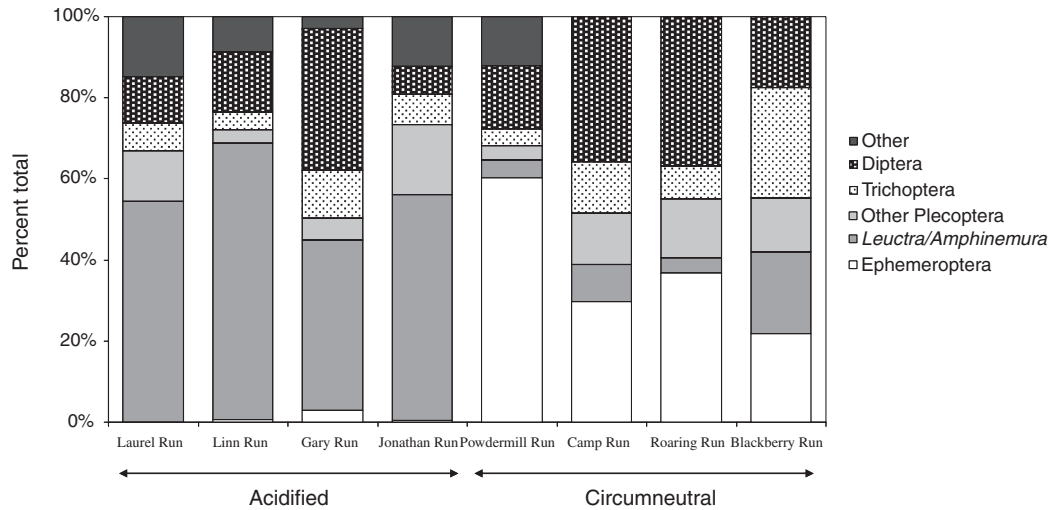


Fig. 3 With moderate levels of stream acidification (4.5–5.5) acid-intolerant mayflies were replaced by two acid-tolerant stonefly genera, *Leuctra* and *Amphinemura*.

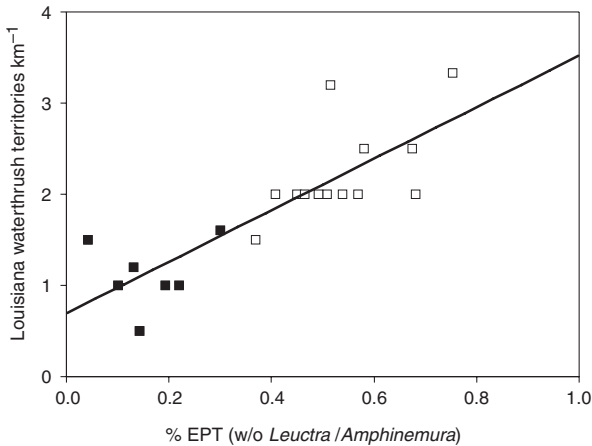


Fig. 4 Breeding density of Louisiana waterthrush on acidified (solid squares) and circumneutral (open squares) streams was strongly related to %EPT in macroinvertebrate samples when two acid-tolerant stonefly genera were excluded.

was initiated on average 26 April ± 1 day, 9 days earlier than on acidified streams, which averaged 5 May ± 2 days ( $t = 3.99, n = 37, P < 0.01$ ). Laying dates for known first nesting attempts tended to be earlier on circumneutral streams (3 May ± 1 day) compared to acidified streams (5 May ± 1 day;  $t = 1.83, n = 153, P = 0.07$ ).

Louisiana waterthrush typically lay five eggs (range 3–6). Average clutch size was smaller on acidified streams compared to circumneutral streams ( $t = -2.03, n = 189, P = 0.04$ ; Table 2). On acidified streams no females laid clutches of six eggs compared

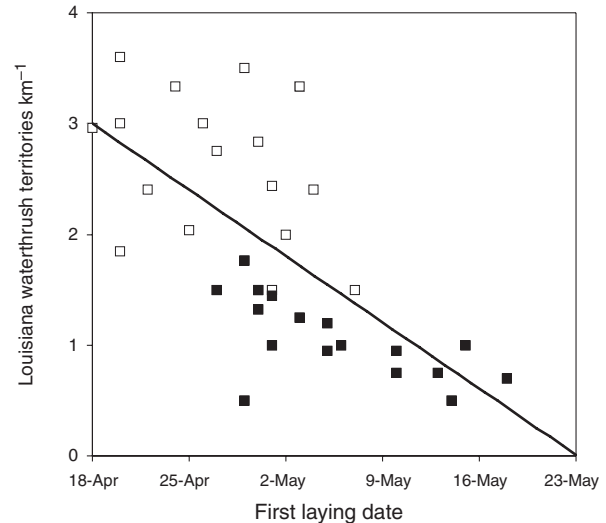


Fig. 5 Louisiana waterthrush breeding density was significantly related to first laying dates on each stream in southwestern Pennsylvania, 1996–2005. The first nests on acidified streams (solid squares) averaged 9 days later than on circumneutral streams (open squares).

to 6% of females on circumneutral streams. However, due to some combination of hatching failure and partial predation of clutches, there was no difference in mean brood size between nests on acidified and circumneutral streams ( $t = -0.61, n = 130, P = 0.53$ ). Overall nest success varied among years (range 23–80%), but low breeding density on acidified streams did not allow us to compare success by year. When combined across years, nesting success did not



**Table 2** Louisiana waterthrush breeding on acidified streams in southwestern Pennsylvania 1996–2005 showed no change in nest success or annual fecundity

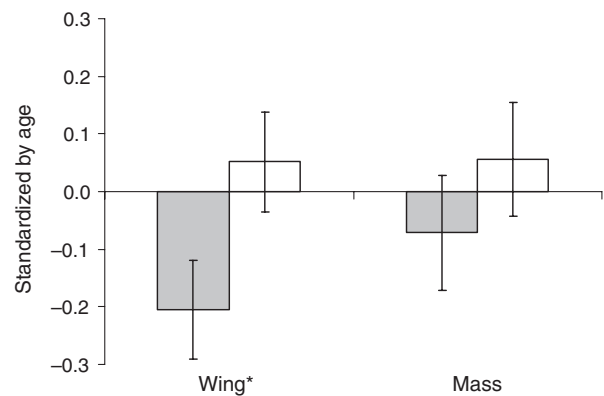
	Acidified	Circumneutral
Daily survival estimate $\pm$ SE	0.980 $\pm$ 0.005	0.973 $\pm$ 0.003
Mayfield success (nests)	57% (46)	49% (146)
Exposure days	738	2189
Clutch size $\pm$ SE (nests)	4.5 $\pm$ 0.1 (40)	4.8 $\pm$ 0.1 (140)*
Brood size $\pm$ SE (nests)	4.2 $\pm$ 0.1 (33)	4.2 $\pm$ 0.1 (102)
Fecundity $\pm$ SE (territories)	3.2 $\pm$ 0.3 (43)	3.4 $\pm$ 0.2 (132)
Fledglings km <sup>-1</sup> $\pm$ SE	2.6 $\pm$ 0.4	5.0 $\pm$ 0.4**

However, females laid smaller clutches and fewer young were fledged on acidified streams (\* $P < 0.05$ ; \*\* $P < 0.01$ ).

differ between acidified and circumneutral streams ( $\chi^2 = 1.44$ , d.f. = 1,  $P = 0.23$ ; Table 2), and fecundity was also not significantly different ( $t = -0.603$ ,  $n = 176$ ,  $P = 0.54$ ; Table 2). However, because of significantly lower breeding density, the number of young fledged km<sup>-1</sup> was significantly lower on acidified streams than on circumneutral streams ( $t = -4.36$ ,  $n = 37$ ,  $P < 0.01$ ; Table 2).

We found no evidence that Louisiana waterthrush adults breeding on acidified streams were smaller or in poorer condition than on circumneutral streams. There was no difference in wing chord or body mass for either males ( $n = 153$ ) or females ( $n = 99$ ) nesting on acidified streams compared to circumneutral streams (all  $t$ -tests:  $P > 0.05$ ). Standardized wing chord lengths of nestlings were significantly shorter on acidified streams than on circumneutral streams ( $t = -2.52$ ,  $n = 458$ ,  $P = 0.01$ ), however, there was no difference in standardized body mass ( $t = -1.12$ ,  $n = 471$ ,  $P = 0.26$ ; Fig. 6).

The proportion of colour-ringed birds known to be site-faithful and/or experienced breeders was marginally different between acidified and circumneutral streams (Fisher's exact:  $P = 0.07$ ; Table 3). Acidified streams had a higher proportion of inexperienced recruits (second-year birds), and many fewer birds nesting on acidified streams returned to breed in multiple years. In contrast, circumneutral streams had fewer recruits overall, a smaller proportion of inexperienced recruits, and a higher proportion of birds returning to breed multiple years (Table 3). Thus,

**Fig. 6** Mean ( $\pm$  SE) standardized wing chord and body mass for 7- to 10-day old nestling Louisiana waterthrush on acidified ( $n = 112$ , gray bars) and circumneutral streams ( $n = 359$ , white bars) in southwestern Pennsylvania, 1996–2005. Standardized wing chords were significantly shorter on acidified streams than on circumneutral streams (\* $P < 0.05$ ).**Table 3** Percentages of color-ringed Louisiana waterthrush on acidified ( $n = 53$ ) and circumneutral ( $n = 148$ ) streams in southwestern Pennsylvania, 1996–2005

	Acidified (%)		Circumneutral (%)	
	Males	Females	Males	Females
Inexperienced Recruits	15	18	10	11
Experienced Recruits	18	4	12	11
Returned Once	18	18	18	14
Returned Again	6	3	16	8

On acidified streams more inexperienced breeders (known age < 1 year old) were recruited into the population compared to experienced birds (known age > 1 year old), and fewer birds returned to breed over multiple years.

acidified streams were characterized by a younger, less site-faithful population compared to circumneutral streams. Several individuals were observed to disperse between study streams. Four colour-ringed birds emigrated from territories on acidified streams to territories on a circumneutral stream between years. A single bird dispersed from one circumneutral stream to another between years, and one bird moved from a circumneutral stream to an acidified stream in the same year, after losing its original territory following a territorial dispute early in the season.

There were marginally significant differences in pairing success between acidified and circumneutral streams (Fisher's exact test:  $P = 0.07$ ). Across all acidified streams and all years, 84% of males

( $n = 55$ ) were paired, while on circumneutral streams 92% ( $n = 152$ ) of males were paired. Unpaired males may contribute to a 'floating' or non-territorial population. Evidence for a population of 'floaters' comes from the bird-ringing programme at PNR where, from 1996 to 2005, 65% ( $n = 29$ ) of waterthrushes captured in the ringing station's suboptimal, non-forested habitat in May were male, and 65% ( $n = 17$ ) of these were young (second-year) birds (R.S. Mulvihill, unpubl. data).

## Discussion

Multiple measures, including lower breeding density, later first laying dates, lower site fidelity, an increased proportion of inexperienced breeders, and opportunistic movements from acidified to circumneutral streams all indicate that stream acidification negatively affects Louisiana waterthrush during the breeding season. Birds appear to occupy acidified streams only as secondary or suboptimal habitat when higher quality habitat is not available. Our results suggest that older or more experienced individuals may be excluding birds from circumneutral streams, consistent with the 'ideal-despotic' model of habitat selection (Fretwell, 1972).

Unexpectedly, we found no difference in reproductive success, daily nest survival estimates, or fecundity between pairs nesting on acidified and circumneutral streams. These observations tend to fit the 'ideal-free' model of habitat selection in which distribution occurs according to resource availability (Fretwell & Lucas, 1970). By decreasing breeding density, pairs appeared to be able to successfully raise young on acidified streams. Other studies have found similar increases in the number of inexperienced young birds in poor quality habitat where the breeding density may be lower and territories larger (Holmes, Marra & Sherry, 1996; Petit & Petit, 1996). In contrast, reduced productivity has been found in dippers breeding on acidified streams (Vickery, 1992). However, in our study, there could be additional costs to adults or young, such as reduced post-breeding survival, which we were unable to detect.

The fact that Louisiana waterthrushes can achieve reproductive success on acidic streams indicates that they are able to compensate for the reduced pH and its associated effects on food availability. One way they are able to compensate is by increasing territory

size. The length of linear territories of several other riparian bird species has also been related to food availability (Davis, 1982; Vickery, 1991; Feck & Hall, 2004). Presumably with reduced food resources, a larger territory is needed to provide sufficient resources for maintenance and reproduction.

However, waterthrush pairs defending longer territories on acidified streams may incur additional costs. We did not test whether adults had to fly farther to obtain food, although given the size of their territories this would seem likely. Birds are likely to forage farther away from the nest when food is less available (Tremblay *et al.*, 2005), and dippers have been shown to spend more time foraging on acidified streams than on circumneutral streams (O'Halloran *et al.*, 1990). Formal and informal surveys of the conspicuous droppings of waterthrush, easily visible on exposed rocks and debris within the wetted stream channel, found much greater concentrations of droppings on circumneutral peripheral tributaries within the territories of birds nesting on acidified streams (Mulvihill, 1999). In addition, during protracted nest watches, we observed adults feeding young novel prey items (e.g. terrestrial salamanders) much more frequently at nests on acidified streams than on circumneutral streams (Mulvihill, 1999). Thus, waterthrushes nesting on acidified streams likely compensated for food shortages associated with lower quality habitat by foraging farther from their nests and by feeding their young larger prey items (see also Tremblay *et al.*, 2005).

Our finding that fecundity (i.e. number of fledglings per pair) did not vary significantly between Louisiana waterthrush pairs on acidified and circumneutral streams was somewhat surprising and may be due primarily to higher rates of nest predation on circumneutral streams. Along with waterthrushes, a variety of potential nest predators co-inhabit headwater streams and associated hardwood forests. These include species like raccoons (*Procyon lotor* L.) and mink (*Neovison vison* Schreber) that, like waterthrushes, also rely to some degree on aquatic prey, as well as corvids, raptors, squirrels and other rodents. Although we did not survey for them, such predators may be less abundant along acidified streams due to reduced prey availability. Lower rates of nest loss in acidified streams compared to circumneutral streams may also be due in part to the greater dispersion of waterthrush nests in these sites (Martin, 1993). If

typical nest predators like corvids and rodents learn to search for nesting birds in systems where nests are more abundant, Louisiana waterthrush pairs nesting at higher densities along circumneutral streams may suffer an increased cost in terms of predation. This density-dependent pattern has been shown in numerous other studies (Stephens & Krebs, 1986; Kamil, Krebs & Pulliam, 1987).

Louisiana waterthrush depend predominantly on aquatic macroinvertebrates for food, both on their breeding and wintering grounds (Robinson, 1995). Our results suggest that reduced stream pH affects Louisiana waterthrush breeding ecology by altering the aquatic macroinvertebrate community. Although overall macroinvertebrate abundance and biomass were lower on some acidified streams, the most consistent difference was that acid-sensitive taxa, such as mayflies, almost completely disappeared from acidified streams and were replaced by two acid-tolerant genera of stoneflies, *Leuctra* and *Amphinemura*. This shift in the macroinvertebrate community has been previously documented to occur at similar pH-levels in the region (Kimmel *et al.*, 1985; Sharpe *et al.*, 1987; Griffith, Perry & Perry, 1995). In our study, density of breeding waterthrush was highly correlated with %EPT but only when these two acid-tolerant stoneflies were excluded from the analyses. Other studies have found that the presence of Louisiana waterthrush is correlated with %EPT (Stucker, 2000; Mattsson & Cooper, 2006), but our results further suggest that mayflies, in particular, may be especially important in the diet of breeding Louisiana waterthrush. Several dipper species (*Cinclus* spp.) that occupy a similar niche world-wide have also been found to consume a high percentage of mayflies when they are available (Ormerod & Tyler, 1991a), and density of breeding dippers has been correlated with favoured mayfly species (Feck & Hall, 2004).

Our results suggest that the effects of stream acidification on Louisiana waterthrush could be most directly linked to food shortages, especially of preferred prey. Results such as the smaller clutch size, as well as smaller nestlings, may be related to abundance of preferred, perhaps higher quality food resources (e.g. Arcese & Smith, 1988; Ormerod & Tyler, 1991a). However, similar effects can occur through calcium shortages (Tilgar, Mänd & Mägi, 2002). Acid precipitation has been associated with reduced calcium availability in soils and decreased abundance and

lower nutritional value of many invertebrate prey species at pH levels below *c.* 6.5 (Scheuhammer, 1991). Crustaceans, mollusks or fish fry are likely the only source of calcium for egg-forming passerines along rivers, since they are unable to gain sufficient calcium when laying daily and feeding on insects (Ormerod & Rundle, 1998). For example, grey wagtail (*Motacilla cinerea* Tunstall) uses mollusks gleaned in the riparian zone, and this may be one of the key differences in flexible foraging that distinguishes its acid-tolerance from *Cinclus*. In another study, calcium supplementation resulted in swallow nestlings with longer feathers prior to fledging (Dawson & Bidwell, 2005).

Environmental changes associated with acid precipitation may reduce calcium in the environment, but can also affect breeding waterthrush through a potential increase in other toxic chemicals (Scheuhammer, 1991; Graveland, 1998; Hames *et al.*, 2002). In dippers, thinner egg shells and decreased levels of calcium in blood serum have been found in acidified areas (Ormerod *et al.*, 1988, 1991). Although we did not examine egg shell thickness or blood serum in Louisiana waterthrush, we found no evidence that adults were in poorer condition on acidified streams. Recent research suggests that waterthrush may be at risk from the effects of elevated levels of mercury (Evers *et al.*, 2005), but moderately elevated levels of mercury do not appear to affect reproduction in dippers (Henny *et al.*, 2005). In addition, some streams in this study, such as Camp Run, recorded slightly elevated levels of aluminum (Table 1). Whether aluminum levels have some responsibility for the consistently lower density of breeding waterthrush on this stream is unknown, but should be studied.

Our results show that Louisiana waterthrush breeding density tends to be lower on acidified streams, and reproductive output per unit area is very different between the two stream categories. In our study area it takes almost double the length of an acidified stream to produce the same number of fledglings as a circumneutral stream. The average density of 2.6 territories km<sup>-1</sup> observed for circumneutral streams in our study is similar to that found elsewhere in the northeastern U.S. (Robinson, 1995). Because of differences in breeding density related to stream pH, Louisiana waterthrush may provide an important indicator of stream acidification, similar to the white-throated dipper (Ormerod *et al.*, 1986; Vickery, 1991;

Buckton *et al.*, 1998). Our results support continued monitoring of this species as an indicator of water quality and biotic integrity of forested headwater streams (O'Connell *et al.*, 2003; Mattsson & Cooper, 2006). Louisiana waterthrush also occupy headwater streams in the Caribbean, Mexico, and Central America during the non-breeding season, where preliminary studies indicate that winter densities also may be influenced by habitat quality (Master *et al.*, 2003). Further work should examine the effects of stream habitat quality on the Louisiana waterthrush throughout its annual cycle.

### Acknowledgments

We thank multiple broods of field assistants, student interns and volunteers for help collecting many of the observations and data used in this study. S.J. Ormerod, C. Townsend and an anonymous reviewer provided helpful comments on an earlier version of this paper. Access to study sites and other logistical support were provided by the Loyalhanna Watershed Association, PA Bureau of Forestry, PA Bureau of State Parks, PA Fish and Boat Commission, PA Bureau of Abandoned Mine Reclamation, Westmoreland and Somerset Conservation District Offices, California University of PA, Saint Vincent College and Penn State University's Environmental Resources Research Institute. Funding for this study was provided by the PA Wild Resource Conservation Program, the U.S. Environmental Protection Agency, the William H. and Ingrid S. Rea and M. Graham Netting research funds of Carnegie Museum of Natural History, the PA Department of Environmental Protection, USDA Natural Resource Conservation Service, Western PA Coalition for Abandoned Mine Reclamation, and the M. Graham and Jane S. Netting Environmental Fund of The Pittsburgh Foundation.

### References

Arcese P. & Smith J.N.M. (1988) Effects of population density and supplemental food on reproduction in song sparrows. *Journal of Animal Ecology*, **57**, 119–136.  
 Buckton S.T., Brewin P.A., Lewis A., Stevens P. & Ormerod S.J. (1998) The distribution of dippers, *Cinclus cinclus* (L.), in the acid-sensitive region of Wales, 1984–1995. *Freshwater Biology*, **39**, 387–396.

Courtney L.A. & Clements W.H. (1998) Effects of acidic pH on benthic macroinvertebrate communities in stream microcosms. *Hydrobiologia*, **379**, 133–145.  
 Davis W.J. (1982) Territory size in *Megaceryle alcyon* along a stream habitat. *Auk*, **99**, 353–362.  
 Dawson R.D. & Bidwell M.T. (2005) Dietary calcium limits size and growth of nestling tree swallows (*Tachycineta bicolor*) in a non-acidified landscape. *Journal of Avian Biology*, **36**, 127–134.  
 Earle J. & Callaghan T. (1998) Impacts of mine drainage on aquatic life, water uses, and man made structures. In: *Coal Mine Drainage Prediction and Pollution Prevention in Pennsylvania* (Eds B.C. Brady, T. Kania, W.M. Smith & R.J. Hornberger), pp. 4.1–4.10. PA Department of Environmental Protection, Harrisburg, PA.  
 Eaton S.W. (1958) A life history study of the Louisiana waterthrush. *Wilson Bulletin*, **70**, 210–235.  
 Evers D.C., Burgess N.M., Champoux L., Hoskins B., Major A., Goodale W.M., Taylor R.J., Poppenga R. & Daigle T. (2005) Patterns and interpretation of mercury exposure in freshwater avian communities in northeastern North America. *Ecotoxicology*, **14**, 193–221.  
 Feck J. & Hall R. (2004) Response of American dippers (*Cinclus mexicanus*) to variation in stream water quality. *Freshwater Biology*, **49**, 1123–1137.  
 Freeman M.C., Pringle C.M. & Jackson C.R. (2007) Hydrologic connectivity and the contribution of stream headwaters to ecological integrity at regional scales. *Journal of the America Water Resources Association*, **43**, 5–14.  
 Fretwell S.D. (1972) *Populations in a Seasonal Environment*. Princeton University Press, Princeton, NJ.  
 Fretwell S.D. & Lucas H.L. (1970) On territorial behavior and other factors influencing habitat distribution in birds. I. Theoretical Development. *Acta Biotheoretica*, **19**, 16–36.  
 Gangewere R.J. (1998) Powdermill Nature Reserve tries the natural approach to fighting pollution from abandoned mines. *Carnegie Magazine*, **59**, 8–13.  
 Graveland J. (1998) Effects of acid rain on bird populations. *Environmental Review*, **6**, 41–54.  
 Griffith M.B., Perry S.A. & Perry W.B. (1995) Macroinvertebrate communities in headwater streams impacted by acidic precipitation in the Central Appalachians. *Journal of Environmental Quality*, **24**, 233–238.  
 Guerold F., Boudot J., Jacquemin G., Vein D., Merlet D. & Rouiller J. (2000) Macroinvertebrate community loss as a result of headwater stream acidification in the Vosges mountains (N-E France). *Biodiversity and Conservation*, **9**, 767–783.

- Hames R.S., Rosenberg K.V., Lowe J.D., Barker S.E. & Dhondt A.A. (2002) Adverse effects of acid rain on the distribution of the Wood Thrush *Hyllocichla mustelina* in North America. *PNAS*, **99**, 11235–11240.
- Henny C.J., Kaiser J.L., Packard H.A., Grove R.A. & Taft M.R. (2005) Assessing mercury exposure and effects to American dippers in headwater streams near mining sites. *Ecotoxicology*, **14**, 709–725.
- Herlihy A.T., Kaufmann P.R., Mitch M.E. & Brown D.D. (1990) Regional estimates of acid mine drainage impact on streams in the Mid-Atlantic and Southeastern United States. *Water, Air and Soil Pollution*, **50**, 91–107.
- Herlihy A.T., Kaufmann P.R., Church M.R., Wigington P.J. & Webb J.R. (1993) The effect of acidic deposition on streams in the Appalachian Mountain and Piedmont Region of the Mid-Atlantic United States. *Water Resources Research*, **29**, 2687–2703.
- Hines J.E. & Sauer J.R. (1989) *Program CONTRAST - A General Program for the Analysis of Several Survival or Recovery Rate Estimates*. Patuxent Wildlife Research Center, U.S. Fish and Wildlife Service, Laurel, MD.
- Holmes R.T., Marra P.P. & Sherry T.W. (1996) Habitat-specific demography of breeding black-throated blue warblers (*Dendroica caerulescens*): implications for population dynamics. *Journal of Animal Ecology*, **65**, 183–195.
- Johnson D.H. (1979) Estimating nest success: the Mayfield method and an alternative. *Auk*, **96**, 651–661.
- Kamil A.C., Krebs J.R. & Pulliam H.R. (1987) *Foraging Behavior*. Plenum Press, New York and London.
- Kimmel W.G., Murphy D.J., Sharpe W.E. & DeWalle D.R. (1985) Macroinvertebrate community structure and detritus processing rates in two southwestern Pennsylvania streams acidified by atmospheric deposition. *Hydrobiologia*, **124**, 97–102.
- Kowalik R.A., Cooper D.M., Evans C.M. & Ormerod S.J. (2007) Acid episodes retard the biological recovery of upland British streams from acidification. *Global Change Biology*, **13**, 2439–2452.
- Likens G.E., Driscoll C.T. & Buso D.C. (1996) Long-term effects of acid rain: response and recovery of a forest ecosystem. *Science*, **272**, 244–246.
- Martin T.E. (1993) Nest predation among vegetation layers and habitat types: revising the dogmas. *American Naturalist*, **141**, 897–913.
- Master T.L., Mulvihill R.S., Leberman R.C., Sánchez J. & Carman E. (2003) *A Preliminary Study of Riparian Songbirds in Costa Rica, With Emphasis on Wintering Louisiana Waterthrushes*. Proceedings of the 3rd International Partners in Flight Conference. March 2002. USDA Forest Service Gen. Tech. Rep. PSW-GTR-191. Asilomar, CA.
- Mattsson B.J. & Cooper R.J. (2006) Toward cost-effective bioassessment: evaluating how Louisiana waterthrushes and habitat assessments serve as indicators of instream biotic integrity. *Freshwater Biology*, **51**, 1941–1958.
- Mayfield H.F. (1975) Suggestions for calculating nest success. *Wilson Bulletin*, **87**, 456–466.
- McCarty J.P. (2001) Variation in nestling tree swallows across multiple temporal and spatial scales. *Auk*, **118**, 176–190.
- Merritt R.W. & Cummins K.W. (1996) *An Introduction to the Aquatic Insects of North America*. 3rd Kendall/Hunt Publishing Co, Dubuque, IA.
- Mulvihill R.S. (1993) Using wing molt to age passerines. *North American Bird Bander*, **18**, 1–10.
- Mulvihill R.S. (1999) Effects of stream acidification on the breeding biology of an obligate riparian songbird, the Louisiana waterthrush (*Seiurus motacilla*). In: *The Effects of Acidic Deposition on Aquatic Ecosystems in Pennsylvania* (Eds W.E. Sharpe & J.R. Drohan), pp. 51–61. Proceedings of the 1998 PA Acidic Deposition Conference, Vol. 2. Environmental Resources Research Institute, University Park, PA.
- O'Connell T.R., Brooks R.P., Laubscher S.E., Mulvihill R.S. & Master T.E. (2003) *Using Bioindicators to Develop a Calibrated Index of Regional Ecological Integrity for Forested Headwater Ecosystems*. Final Report to U.S. Environmental Protection Agency, STAR Grants Program. Report No. 2003-01. Penn State Cooperative Wetlands Center, Penn State University, University Park, PA.
- O'Halloran J., Gribbin S.D., Tyler S.J. & Ormerod S.J. (1990) The ecology of dippers *Cinclus cinclus* (L.) in relation to stream acidity in upland Wales: time-activity budgets and energy expenditure. *Oecologia*, **85**, 271–280.
- Ormerod S.J. & Rundle S.D. (1998) Effects of experimental acidification and liming on terrestrial invertebrates: implications for calcium availability to vertebrates. *Environmental Pollution*, **103**, 183–191.
- Ormerod S.J. & Tyler S.J. (1991a) Exploitation of prey by a river bird, the dipper *Cinclus cinclus* (L.) along acidic and circumneutral streams in upland Wales. *Freshwater Biology*, **25**, 105–116.
- Ormerod S.J. & Tyler S.J. (1991b) The influence of stream acidification and riparian land use on the feeding ecology of grey wagtails *Motacilla cinerea* in Wales. *Ibis*, **133**, 53–61.
- Ormerod S.J., Allinson N., Hutchinson D. & Tyler S.J. (1986) The distribution of breeding dippers (*Cinclus cinclus* (L.); Aves) in relation to stream acidity in upland Wales. *Freshwater Biology*, **16**, 501–507.
- Ormerod S.J., Bull K.R., Cummins C.P., Tyler S.J. & Vickery J.A. (1988) Egg mass and shell thickness in

- dippers (*Cinclus cinclus*) in relation to stream acidity in Wales and Scotland. *Environmental Pollution*, **55**, 107–121.
- Ormerod S.J., O'Halloran J., Gribbin S.D. & Taylor S.J. (1991) The ecology of dippers *Cinclus cinclus* in relation to stream acidity in upland Wales: breeding performance, calcium physiology, and nestling growth. *Journal of Applied Ecology*, **28**, 419–433.
- Petit L.J. & Petit D.R. (1996) Factors governing habitat selection by prothonotary warblers: field tests of the Fretwell-Lucas models. *Ecological Monographs*, **66**, 367–387.
- Pyle P. (1997) *Identification Guide to North American Birds, Part 1*. Slate Creek Press, Bolinas, CA.
- R Development Core Team (2007) *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna. ISBN 3-900051-07-0. Available at: <http://www.r-project.org> (last accessed on 30 May 2008).
- Rich T.D., Beardmore C.J., Berlanga H. *et al.* (2004) *Partners in Flight North American Landbird Conservation Plan*. Cornell Lab of Ornithology, Ithaca, NY.
- Robinson D.W. (1995) Louisiana waterthrush (*Seiurus motacilla*). In: *The Birds of North America, No. 151* (Eds A. Poole & F. Gill), pp. 1–18. The Academy of Natural Sciences; American Ornithologists' Union, Philadelphia, PA; Washington D.C.
- Rutt G.P., Weatherley N.S. & Ormerod S.J. (1990) Relationships between the physicochemistry and macroinvertebrates of British upland streams: the development of modeling and indicator systems for predicting fauna and detecting acidity. *Freshwater Biology*, **24**, 463–480.
- Scheuhammer A.M. (1991) Effects of acidification on the availability of toxic metals and calcium to wild birds and mammals. *Environmental Pollution*, **71**, 329–375.
- Schindler D.W. (1988) Effects of acid rain on freshwater ecosystems. *Science*, **239**, 149–157.
- Schreiber R.K. & Newman J.R. (1988) Acid precipitation effects on forest habitats: implications for wildlife. *Conservation Biology*, **3**, 249–259.
- Sharpe W.E., Perlic T.G., Tzilkowski W.M. & Kimmel W.G. (1987) Status of headwater benthic insect populations in an area of high hydrogen ion and sulfate deposition. *Northeastern Environmental Science*, **6**, 23–30.
- Stephens D.W. & Krebs J.R. (1986) *Foraging Theory*. Princeton University Press, Princeton, NJ.
- Stucker H.S. (2000) *Biodiversity of Southeastern Minnesota Forested Streams: Relationships Between Trout Habitat Improvement Practices, Riparian Communities and Louisiana Waterthrushes*. Master's Thesis, University of Minnesota, St Paul, MN.
- Tilgar V., Mänd R. & Mägi M. (2002) Calcium shortage as a constraint on reproduction in great tits *Parus major*: a field experiment. *Journal of Avian Biology*, **33**, 407–413.
- Tremblay I., Thomas D., Blondel J., Perret P. & Lambrechts M.M. (2005) The effect of habitat quality on foraging patterns, provisioning rate and nestling growth in Corsican blue tits *Parus caeruleus*. *Ibis*, **147**, 17–24.
- Vickery J. (1991) Breeding density of dippers (*Cinclus cinclus*), grey wagtails (*Motacilla cinerea*) and common sandpipers (*Actitis hypoleucos*) in relation to the acidity of streams in south-west Scotland. *Ibis*, **133**, 178–185.
- Vickery J. (1992) The reproductive success of the dipper *Cinclus cinclus* in relation to the acidity of streams in south-west Scotland. *Freshwater Biology*, **28**, 195–205.

(Manuscript accepted 7 May 2008)





# Freshwater Biology

Volume 53, Number 11, November 2008

## CONTENTS

### Original Articles

- 2133 *Jeff M. Sereda, Jeff J. Hudson and Philip D. McLoughlin*: General empirical models for predicting the release of nutrients by fish, with a comparison between detritivores and non-detritivores
- 2145 *Robin Hale, Barbara J. Downes and Stephen E. Swearer*: Habitat selection as a source of inter-specific differences in recruitment of two diadromous fish species
- 2158 *Robert S. Mulvihill, Felicity L. Newell and Steven C. Latta*: Effects of acidification on the breeding ecology of a stream-dependent songbird, the Louisiana waterthrush (*Seiurus motacilla*)
- 2170 *Pieter Vanormelingen, Karl Cottenie, Erik Michels, Koenraad Muylaert, Wim Vyverman and Luc De Meester*: The relative importance of dispersal and local processes in structuring phytoplankton communities in a set of highly interconnected ponds
- 2184 *Thomas Lakowitz, Christer Brönmark and Per Nystrom*: Tuning in to multiple predators: conflicting demands for shell morphology in a freshwater snail
- 2192 *Timothy D. Jardine, Karen A. Kidd, John T. Polhemus and Richard A. Cunjak*: An elemental and stable isotope assessment of water strider feeding ecology and lipid dynamics: synthesis of laboratory and field studies
- 2206 *Daniel J. McGarvey and G. Milton Ward*: Scale dependence in the species-discharge relationship for fishes of the southeastern U.S.A.
- 2220 *George W. Peck and William E. Walton*: Effect of mosquitofish (*Gambusia affinis*) and sestonic food abundance on the invertebrate community within a constructed treatment wetland
- 2234 *Brian T. Greene, Winsor H. Lowe and Gene E. Likens*: Forest succession and prey availability influence the strength and scale of terrestrial-aquatic linkages in a headwater salamander system
- 2244 *Birgitta Malm Renöfält and Christer Nilsson*: Landscape scale effects of disturbance on riparian vegetation
- 2256 *Wen Xiong, Dan Yu, Qiang Wang, Chunhua Liu and Ligong Wang*: A snail prefers native over exotic freshwater plants: implications for the enemy release hypotheses
- 2264 *Bram Vanschoenwinkel, Aline Waterkeyn, Tim Vandecaetsbeek, Olivier Pineau, Patrick Grillas and Luc Brendonck*: Dispersal of freshwater invertebrates by large terrestrial mammals: a case study with wild boar (*Sus scrofa*) in Mediterranean wetlands
- 2274 *Scott H. Markwith and David S. Leigh*: Subaqueous hydrochory: open-channel hydraulic modelling of non-buoyant seed movement
- 2287 *Juha Jurvelius and Timo J. Marjomäki*: Night, day, sunrise, sunset: do fish under snow and ice recognize the difference?
- 2295 *Dean Jacobsen and Klaus Peter Brodersen*: Are altitudinal limits of equatorial stream insects reflected in their respiratory performance?
- 2309 *Gerald Louette, Luc de Meester and Steven DeClerck*: Assembly of zooplankton communities in newly created ponds
- 2321 *Wei Li, Zhao Zhang and Erik Jeppesen*: The response of *Vallisneria spirulosa* (Hydrocharitaceae) to different loadings of ammonia and nitrate at moderate phosphorus concentration: a mesocosm approach

### Applied Issues

- 2331 *Cassie D. Mellon, Mark S. Wipfli and Judith L. Li*: Effects of forest fire on headwater stream macroinvertebrate communities in eastern Washington, U.S.A.

FRONT COVER: A Louisiana waterthrush poised on a rock at the water's edge. Like dippers and wagtails, the waterthrush feeds extensively on the adult and immature stages of aquatic insects. Like other species that forage at the interface of land and moving water, the waterthrush 'bobs' its tail up and down continually as it forages. Its specific scientific name, *motacilla*, comes from the Greek and means moving or wagging tail – a scientific name that it shares with the true wagtails (Motacillidae). From Robert S. Mulvihill *et al.*, p.2158. The painting is an original watercolour by world renowned bird sculptor, Larry Barth.

Information on this journal can be accessed at <http://www.blackwellpublishing.com/fwb>

This journal is available online at Wiley InterScience. Visit [www.blackwell-synergy.com](http://www.blackwell-synergy.com) to search the articles and register for table of contents e-mail alerts.

